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استخراج و پیش تغلیظ یون جیوه با استفاده از نانوذرات مغناطیسی عاملدارشده قبل از تعیین با دستگاه اسپکتروفتومتر مرئی- ماورائ بنفش

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Extraction and Preconcentration of Trace Amounts of Mercury Using Functionalized Magnetic-Nanoparticles Prior to Their Determination by UV-Vis Spectrophotometry

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چکیدہ

در این مقاله، نانوذرات مغناطیسی عاملدارشده با لیگاند تری آزن برای استخراج و پیش تغلیظ یون جیوه در نمونههای آبی با دستگاه اسپکتروفتومتر مرئی-ماورائ بنفش اندازه گیری شده است. ویژگیهای نانو ذرات اصلاحشده با تکنیکهای مختلف از جمله میکروسکوپ روبش الکترونی، میکروسکوپ الکترونی عبوری و دیفراکسیون اشعه ایکس مورد ارزیابی قرار گرفت. در فرایند جداسازی، محلول آبی یون جیوه با نانوذرات مغناطیسی اکسید آهن اصلاحشده با لیگاند تری آزن مخلوط شد و سپس یک میدان مغناطیسی خارجی برای جداسازی نانو ذرات شامل یون جیوه و میش تفاده قرار گرفت. شرایط آزمایشگاهی برای جذب موثر شامل مقدار اصلاح گر، مقدار نانو ذره و نوع شوینده بررسی گردید. تحت شرایط بهینه استخراج و پیش تغلیظ، دامنه خطی منحنی کالیبراسیون از ۲۰۸–۴ میکرو گرم بر لیتر با ضریب رگراسیون ۲۹۹۲ برای ۲۰ برای اندازه گیری، حد تشخیص ۱/۰۵ میکروگرم بر لیتر و فاکتور پیش تغلیظ ۳۸ بدست آمد. این تکنیک برای نمونهای آبی یون جیوه موفقیت آمیز بوده است.

واژههای کلیدی

ای-۱- (۲- اتوکسی فنیل)-۳- (۴- نیتروفنیل) تری آز ۱- ان؛ یون جیوه؛ ذرات نانومغناطیسی، پیش تغلیظ؛ اسپکترو فتومتری ماورابنفش -مرئی.

Abstract

In this paper, magnetic nanoparticles functionalized by triazene ligand were prepared for extraction/preconcentration of trace amounts of mercury ions in water samples former on its determination with UV-Vis spectrophotometry. The modified magnetic nanoparticles were characterized by the various techniques such as Scanning Electron Microscope (SEM), Transmission Electron Microscopy (TEM) and X –Ray Diffractometry (XRD). In the separation process, aqueous solution of Hg²⁺ ion was mixed with Fe₃O₄ magnetite nanoparticles modified with (E)-1-(2-Ethoxyphenyl)-3-(4-nitrophenyl)triaze-1-ene (ENT), (MNPs@SiO₂-ENT) and then external magnetic field was applied for isolation of magnetite nanoparticles containing mercury ions. Experimental conditions for effective adsorption including pH, ENT amount, Fe₃O₄ NPs amounts, and eluent type have been studied and established. Under the optimal extraction and preconcentration conditions, calibration curve was linear in the range of 4–80 μ gL⁻¹ with r² = 0.9992 (n=20), the limit of detection (LOD) 1.05 μ g L⁻¹ and enrichment factor was 38, respectively. This technique was successfully used for the determination of Hg (II) in aqueous samples.

Keywords

(E)-1-(2-Ethoxyphenyl)-3-(4-nitrophenyl) triaze-1-ene; Hg^{2+} ions; Magnetic Nanoparticles; Preconcentration; UV-Vis Spectrophotometry.

1. INTRODUCTION

Determination of traces of mercury in the environmental samples is of great importance due to its high toxicity, high reactivity, accumulative character, wide usage and relative solubility in water and living tissues [1-5]. Toxic effects include damage to the brain, kidney, and lungs. Mercury pollution constitutes serious worldwide environmental problem [6-9].

Mercury can solve in natural water and change over to methyl mercury, neurotoxic compound that interferes in protein synthesis by blockages of the enzyme binding sites, by the action of aerobic microorganisms [10]. Recent reports evaluate a total mercury concentration in non-polluted water ranging from 0.006 to 5 ng mL⁻¹ [11-16]. Hg (II) can diffuse and accumulate into the aquatic ecosystem as contaminant in the environment. Therefore, determination of Hg (II) levels in water samples is very important. Total mercury concentration in marine species and natural water is about $0.1-200 \text{ ng mL}^{-1}$ and the allowed mercury level set by World Health Organization (WHO) for drinking water is 1 ng mL⁻¹ [17, 18]. Therefore, a great variety of techniques have been applied for the determination of Hg (II) ions in trace levels, such as cold vapor atomic absorption spectrometry(CV-AAS) [19], inductively coupled plasma atomic emission /or mass spectrometry [20,21], spectrofluorimetry [22, 23], atomic florescence spectrometry [24-26] gas chromatography[27] X-ray spectrometry[28,29]. However, these high-performance techniques require expensive instruments and technical expertise in sample preparation steps. Nowadays, the colorimetric method acting is a common technique used for analysis without prerequisite of complex instrumental design and carrying out. The method is based on monitoring of the color change of solutions, e.g. by UV-visible absorption spectrophotometry, which is related to concentration of analytes.

This work concentrates on advancement a simple and sensitive spectrophotometric method to determination of the Hg²⁺ ion. In this work, (MNPs@SiO₂) were prepared and functionalized by (E)-3-(2-ethoxyphenyl)-1-(4-nitrophenyl) triaz-1-ene (ENT), (Scheme 1) to create (MNPs@SiO₂-ENT). Among magnetic nanoparticles, Fe₃O₄ nanoparticles are highly dispersible in solutions. with particle sizes of less than 40 nm. They offer a large surface area and super paramagnetic properties; they are attracted to a magnetic field but do not retain magnetic properties when the field is removed, making them highly useful in novel separation processes [19, 20]. The prepared (MNPs@SiO₂-ENT) was utilized as a novel dispersible solid phase extraction sorbent that can be magnetically collected and has a high capacity and selectivity toward mercury ion. This method is very simple and fast separation that needs a little amount of adsorbent for extraction.



Scheme 1. Preparation steps for fabrication of ENTfunctionalized magnetic Fe₃O₄ nanoparticle

2. EXPERIMENTAL

2.1. Reagents and solutions

All used reagents were of analytical grade. Ferric chloride(FeCl₃•6H₂O),ferrous chloride (FeCl₂•4H₂O), sodium hydroxide, mercurv chloride (HgCl₂), sodium acetate, tetraethyl orthosilicate (TEOS), methanol, acetone. acetonitrile, nitric acid and hydrochloric acid were all purchased from Merck Darmstadt, Germany). The stock solution of Hg (II) (1000 mgL⁻¹) was prepared by dissolving the 0.0135g of HgCl₂ in doubly distilled water and 1 mL H₂SO₄ (0.1 mol L⁻¹) and diluted to 10 mL in a volumetric flask. Working solutions were prepared from the stock solution by serial dilutions with doubly distilled water. A solution of $(1 \times 10^{-4} \text{ mol } \text{L}^{-1})$ triazene ligand was prepared by dissolving the appropriate amount of ENT in 1, 4-Dioxane and was then kept away from the light. A buffer solution (pH 4) was prepared by mixing a sodium acetate solution (0.1mol L⁻¹) with a concentrated hydrochloric acid (12 mol L⁻¹) solution.

2.2. Instrument

UV-Vis Perkin-Elmer (Lambda) PC spectrophotometer with 1 cm cell quartz was used for recording all spectra and absorbance measurements. The pH measurements were conducted by a pH-meter (Metrohm-827) calibrated against two standard buffer solutions. Scanning electron microscopy (SEM) was taken with KYKY, EM-3200, TEM images of the nanoparticle were obtained by using a Zeiss EM 900 transmission electron microscope (Carl Zeiss, Germany). The powder XRD pattern was collected with a Philips-PW 12C diffractometer (Philips PW, Amsterdam, The Netherlands) using Cu Ka radiation.

2.3. Synthesis of Fe₃O₄ nanoparticles (MNPs)

The chemical co-precipitation method was used to preparation of the Fe₃O₄ nanoparticles (MNPs) [21]. First, for preparation of a stock solution, FeCl₃•6H₂O (0.037 mol, 9.94 gr) with (0.0081 mol, 1.62 g) of FeCl₂•4H₂O and in 10.0 mL of HCl (2 mol L⁻¹) were mixed in a beaker which was then degassed using nitrogen gas for 20 min before use. Simultaneously, 100 mL of (3 mol L⁻ ¹) NaOH solutions was degassed (for 15 min) and heated to 70°C in a reactor. The stock solution was then added drop wise using a dropping funnel for 30 min under nitrogen gas protection and vigorously stirring (1000 rpm) using a glassware stirrer. During the whole process, the solution temperature was maintained at 65-70°C and nitrogen gas was used to prevent the intrusion of oxygen. The color of the solution changed immediately from orange to dark after addition of NaOH solutions. The pH of the suspension was adjusted to pH=7.0 using HCl (2 mol L⁻¹). The magnetic nanoparticles were separated from the resulting solution with a magnet and the supernatant was discarded. The obtained MNPs were rinsed twice with deionized water and ethanol. Finally, MNPs were dried in oven at 50°C for 4 hours. The obtained MNPs were stable under these conditions for up to about one month.

2.4. Preparation of silica coated (MNPs@SiO₂)

The Fe₃O₄@SiO₂ Nanospheres were prepared by a modified Stöber method [30]. Briefly, MNPs (0.50 g) was dispersed in a mixture of ethanol (50 mL), deionized water (5 mL) and TEOS (0.20 mL), followed by the addition of 5.0 mL of NaOH (10 wt.%). This solution was stirred mechanically for 30 min at room temperature. Then the product, MNPs@SiO₂, was separated by an external magnet and was washed with deionized water and ethanol three times and dried at 80°C for 10 h.

2.5. Preparation of ENT functionalized MNPs (MNPs@SiO₂-ENT)

For the preparation of ENT modified magnetic nanoparticles, triazene ligand (solution in acetonitrile, 1 mL, and 1.3×10^{-3} mol L⁻¹) was added to 1 gr of MNPs@SiO₂ and stirred for 1h in ultrasonic bath. Subsequently, the suspension was stood for half one hour before separating and washing with ethanol. After five cycles of separation/washing/re-dispersion with ethanol, the powder obtained was oven dried at 50 °C for 10 hours.

2.6. Characterization of MNPs@SiO₂-ENT

In order to investigate the morphology and particle size of the prepared MNPs@SiO₂-ENT

scanning electron microscopy (SEM) was performed and results were depicted in Fig. 1. The SEM image of MNPs@SiO₂-ENT shows spherical and uniform nanoparticles with size range of <100 nm. As shown in Fig. 3, during the modification of MNPs@SiO₂, a considerable change was not taken place on shape and size of nanoparticles but partial agglomeration of nanoparticles was observed.

XRD pattern of magnetite nanoparticles is shown in Fig. 2. Both Fe₃O₄ and MNPs@SiO₂-ENT show diffraction peaks at 2θ = 30.3, 35.6, 43.3, 53.8, 57.4 and 62.9 that are indexed to the crystalline cubic inverse spinel structure of Fe₃O₄ nanoparticles.

Fig. 4. Shows the magnetization versus applied field of MNPs@SiO₂-ENT that was obtained by VSM. The saturation magnetization value was measured to be ~60 emu g⁻¹ for Fe₃O₄ and ~40 emu g⁻¹ for MNPs@SiO₂-ENT. The results show that surface modification of MNPs has insignificance effect on the magnetic properties of MNPs.



26 KV 40.0 KX 1 um KYKY-EM3200 SN:0589 Fig. 1. SEM image of MNPs@SiO₂-ENT.



Fig. 2. X-ray diffraction patterns of a) MNPs and b) $MNPs@SiO_2$ -ENT.



Fig. 3. TEM image of MNPs@SiO₂-ENT.



Fig. 4. Vibrational of sample MNPs@SiO₂-ENT vs. magnetization curve.

2.7. Absorption of Hg^{2+}

The procedure for the magnetic solid-phase extraction of Hg²⁺ ion in a batch experiment is as follows: A solid adsorbent (1 gr) was added to a 50 mL of Hg²⁺ ion solution (100 μ g L⁻¹), then the pH value was adjusted to 4.0 with acetate/ acetic buffer and the solution was shaken in ultrasonic batch for 10 min to facilitate adsorption of the metal ions onto the nanoparticles. Then the magnetic adsorbents was separated easily and quickly using an external magnet and decanted directly. Subsequently, 2 mL of 1 mol L⁻¹ HNO₃ solution was added as eluent. Finally, the magnet was used again to settle the magnetic nanoparticles and elute was separated for spectrophotometric analysis. In this section, we used change color dithizone for determination of Hg²⁺ ion. The amount of Hg²⁺ adsorbed onto the absorbent was calculated using the concentration of before and after absorption as follows $X = (C_0-$ C) V/M where X is amounts of Hg^{2+} adsorbed, Hg^{2+} (mg/g), C_0 is the initial concentration before adsorption (mg L⁻¹), C is the concentration after absorption (mgL⁻¹), V is the solution volume (L), and M is the mass of adsorbent.

3. RESULT AND DISCUSSION

3.1. Effect of pH

The pH of the sample solution plays a key role in the solid phase extraction procedure. A proper pH value can improve the adsorption efficiency, and diminish interference from the medium. The effect of pH on the extraction mercury was studied between the pH ranges of 2–7, the results are shown in Fig. 6. Based on the obtained results, the extraction of Hg (II), increased as the pH in the aqueous solution was increased from 2 to 4, and then it decreased when the pH increased from 4 to 6. Therefore, the optimum pH was 4 for extraction of Hg (II) from the aqueous solution. At strong acidic media, pH (<2) the nitrogen atoms of ligand could be protonated, so the stability of complex formation between the sorbent and Hg (II) ions were reduced. It is worthy to note that no precipitation of the mercury hydroxide is expected in the pH range of 4-7 [31]. The higher pH values (>9) were not studied because functionalized Fe₃O₄ magnetic nanoparticles were not stable in alkali solutions due to the breaking of the Si-O-Si bonds by hydroxide ion attack.



Fig. 5. Absorbance spectra of Dithizone solution $(5.0 \times 10^{-5} \text{ mol } \text{L}^{-1})$ in 1, 4-Dioxane and increasing concentration of Hg(II) ion solution $(1.30 \times 10^{-3} \text{ mol } \text{L}^{-1})$. The arrows show the direction of absorbance by increasing the ion concentration.



Fig. 6. Effect of pH on the extraction of Hg(II) ions.

3.2. Effect of ligand amount

Chemical structure of ENT ligand is shown in Fig. 7. ENT can form Hg-ENT complexes with stoichiometry's of 1:2 (Hg: ENT) [32]. The advantages of using ENT can be summarized as follows: (a) ENT forms selective complex with Hg (II). In addition, Hg (II) gives a color complex with ENT, it is especially suitable when the detection system used is a spectrophotometer; (b) reaction between Hg (II) and ENT is extremely fast. In order to investigate the effect of the amount of ENT on the quantitative extraction of Hg (II), the extraction was conducted by varying the amounts of ENT from 143 to 715 mg and the results are shown in Fig. 8. As can be seen, the obtained results showed that the extraction of Hg (II) ion was quantitative by using only 429 mg of ENT. Subsequent extraction experiments were carried out with 429 mg of ENT in order to achieve high capacity and to account for other extractable species.



Fig. 7. Chemical structure of the Triazene ligand.



Fig. 8. Effect of ligand amount on the extraction of Hg(II) ions.

3.3. Effect of MNPs@SiO₂-ENT dosage

In order to investigate the effect of MNPs@SiO₂-ENT amount on the quantitative extraction of Hg (II) ions, the extraction was conducted by varying amount of the modified MNPs@SiO₂-ENT from 100 to 800 mg. The obtained results showed that the extraction of Hg (II) was quantitative by using only 400 mg of MNPs@SiO₂-ENT. Subsequent extraction experiments were carried out with 400 mg of MNPs@SiO₂-ENT in order to achieve higher capacity and to account for other extractable species.

3.4. Choice of eluent

Some experiments were carried out in order to choose the proper eluent for the adsorbed Hg (II)

ions after the extraction of 100 μ g L⁻¹ of Hg (II) in a 50 mL solution by 400 mg MNPs@SiO₂-ENT. The Hg (II) ions were stripped with different volumes of various acids including HNO₃, HCl, HClO₄ and CH₃COOH, which could accomplish the quantitative elution of Hg (II) from the modified MNPs. It is obvious from the data given in Fig. 10, 11 that 2.0 mL of 1 mol L⁻¹ HNO₃ gave the best results for preconcentration factor.



Fig. 9. Effect of the adsorbent amount on the extraction of Hg(II) ions.



Fig. 10. Choice of eluent: The Hg(II) ions were stripped with different of eluent.



Fig. 11. Volume of eluent: The Hg(II) ions were stripped with different of volume.

3.5. Analytical performance

The method showed a good linearity over the calibration range $(4-80 \ \mu g \ L^{-1})$ and the correlation of determination (r^2) was higher than 0.9992. The

limit of detection (LOD = $1.05 \ \mu g \ L^{-1}$) was calculated as $3S_b/m$ (S_b , standard deviation of the blank signals and m, slope of calibration curve after preconcentration). An enhancement factor of 38 was calculated as the ratio between the slopes of the calibration curve submitted by the SPE procedure (for V = 50 mL) and the direct calibration curve (without preconcentration).

4. CONCLUSION

A novel magnetic adsorbent of silica coated Fe₃O₄ magnetic nanoparticle that modified with triazene ligand (MNPs@SiO2-ENT) was prepared and applied for preconcentration and determination of Hg (II) ions from water samples by using dithizone as a reagent and UV-Vis spectrophotometry. The proposed method offers a simple, sensitive, and inexpensive method for extraction and determination of Hg(II). Furthermore, Hg(II) was quantitatively extracted and an enhancement factor of 38-fold was obtained. The results demonstrated that the method is very suitable for the rapid extraction of Hg(II) ions from large volumes of water samples at ppb levels besides having both good accuracy and precision. In addition, sensitivity of the method is enough for the determination of Hg (II) in variety of environmental samples. The proposed method can be a guideline for the extraction of other metal ions from large volumes of sample.

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REFRENCESE

[1] D.K. Singh and S. Mishra, Synthesis and characterization of Hg (II)-ion-imprinted polymer: kinetic and isotherm studies, *Desalin* 257 (2010) 177-183.

[2] L. Mergola, S. Scorrano, E. Bloise and M. Pia Di Bello, Novel polymeric sorbents based on imprinted Hg (II)-diphenylcarbazone complexes for mercury removal from drinking water, *Polymer J.* 48 (2016) 73-79.

[3] N.S. Khairi, NA. Yusof, A.H. Abdullah and F. Mohammad, Removal of Toxic Mercury from Petroleum Oil by Newly Synthesized Molecularly-Imprinted Polymer, *J. Mol. Sci.* 16 (2015) 10562 -10577.

[4] H.R. Rajabi, M. Roushani and M. Shamsipur, Development of a highly selective voltammetric sensor for nanomolar detection of mercury ions using glassy carbon electrode modified with a novel ion imprinted polymeric nanobeads and multi-wall carbon nanotubes, *J. Electroanal. Chem.* 693 (2013) 16-22.

[5] A. Afkhami, T. Madrakian, S.J. Sabounchei, M. Rezaei, S. Samiee and M. Pourshahbaz, Construction of a modified carbon paste electrode for the highly selective simultaneous electrochemical determination of trace amounts of mercury (II) and cadmium (II), *Sensor. Actuat. B-Chem.* 161 (2012) 542 -548.

[6] A. Shokrollahi, M. Ghaedi and M. Shamsipur, Highly selective transport of mercury (II) ion through a bulk liquid membrane, *Quim. Nova* 32 (2009) 153-157.

[7] I. Ghodbane and O. Hamdaoui, Removal of mercury (II) from aqueous media using eucalyptus bark: Kinetic and equilibrium studies, J. *Hazard. Mater*, 160 (2008) 301-309.

[8] M. Ghaedi, M.R. Fathi, A. Shokrollahi and F. Shajarat, Highly selective and sensitive preconcentration of mercury ion and determination by cold vapour atomic absorption spectroscopy *Anal. Lett.* 39 (2006) 1171-1185.

[9] D.M. Manohar, K.A. Krishnan and T.S. Anirudhan, Removal of mercury (II) from aqueous solutions and chlor-alkali industry wastewater using 2-mercaptobenzimidazole-clay, *Water. Res.*, 36 (2002) 1609-1619.

[10] M. Tuzen, I. Karaman, D. Citak and M. Soylak, Mercury (II) and methyl mercury determinations in water and fish samples by using solid phase extraction and cold vapour atomic absorption spectrometry combination, *Food. Chem. Toxicol*, 47 (2009)1648-1652.

[11] C. Huang, B. Hu, Silica-coated magnetic nanoparticles modified with γ -mercaptopropyltrimethoxysilane for fast and selective solid phase extraction of trace amounts of Cd, Cu, Hg, and Pb in environmental and biological samples prior to their determination by inductively coupled plasma mass spectrometry, *Spectrochim. Acta B.* 63 (2008) 437-444.

[12] M. Baghdadi and F. Shemirani, Coldinduced aggregation micro extraction: a novel sample preparation technique based on ionic liquids, *Anal. Chim. Acta.* 613 (2008) 56-63.

[13] Z. Li, Q. Wei, R. Yuan, X. Zhou, H. Liu, H. Shan and Q. Song, A new room temperature ionic liquid 1-butyl-3-trimethylsilylimidazolium hexa-fluorophosphate as a solvent for extraction and preconcentration of mercury with determination by cold vapor atomic absorption, spectrometry, *Talanta* 71 (2007) 68-72.

[14] N. Pourreza and K. Ghanemi, Determination of mercury in water and fish samples by cold vapor atomic absorption spectrometry after solid phase extraction on agar modified with 2-mercaptobenzimidazole, *j. Hazard. Mater* 161 (2009) 982-987.

[15] R. Ito, M. Kawaguchi, N. Sakui, H. Honda, N. Okanouchi, K. Saito and H. Nakazawa, Mercury speciation and analysis in drinking water by stir bar sorptive extraction with in situ propyl derivatization and thermal desorption–gas chromatography–mass spectrometry, *J. Chromatogr A* 1209 (2008) 267-270.

[16] J.C.A. de Wuilloud, R.G. Wuilloud, M.F. Silva, R.A. Olsina and L.D. Martinez, Sensitive determination of mercury in tap water by cloud point extraction pre-concentration and flow injection-cold vapor-inductively coupled plasma optical emission spectrometry, *Spectrochim. Acta B.* 57 (2002) 365-374.

[17] R. Eisler, Mercury Hazards to Living Organisms, 1st ed, *Taylor & Francis* (2006) 155. [18] L. Ebdon, M.E. Foulkes, S.L. Roux and R. Munoz-Olivas, Cold vapour atomic fluorescence spectrometry and gas chromatography-pyrolysisatomic fluorescence spectrometry for routine determination of total and organometallic mercury in food samples, *Analyst*, 127 (2002) 1108-1114.

[19] A. Detcheva and K.H. Grobecker, Determination of Hg, Cd, Mn, Pb and Sn in seafood by solid sampling Zeeman atomic absorption spectrometry, *Spectrochim. Acta B* 61, (2006) 454-459.

[20] K. Leopold, M. Foulkes and P.J. Worsfold, Preconcentration techniques for the determination of mercury species in natural waters *TrAC*, *Trends. Anal. Chem.* 28 (2009) 426-435.

[21] O.T. Butler, M. Cook, C.F. Harrington, S.J. Hill, J. Rieuwerts and D.L. Miles, Atomic spectrometry update. Environmental analysis, J. *Anal. Atom. Spectrom.* 22 (2007) 187-221.

[22] C. Huang, and B. Hu, Speciation of inorganic tellurium from seawater by ICP-MS following magnetic SPE separation and preconcentration, *J. Sep. Sci.* 31 (2008) 760-767.

[23] M. Shamsipur, M. Hosseini, K. Alizadeh, N. Alizadeh, A. Yari, C. Caltagirone and V. Lippolis, Novel fluorimetric bulk optode membrane based on a dansylamidopropyl pendant arm derivative of 1-aza-4, 10-dithia-7-oxacyclododecane ([12] aneNS₂O) for selective subnanomolar detection of Hg(II) ions, *Anal. Chim. Acta.*533 (2005) 17-24.

[24] G. Vasile, N.M. Marin, L.V. Cruceru, M. Simion, T. Galaon, andC.B. Lehr, Determination of ultra-trace mercury in water samples based on cold vapor atomic fluorescence spectrometry using a gold trap, *Rev. Chim.* 67 (2016) 1427–1432.

[25] O.Yayayuruk, E. Henden, Use of Ni/NixB nanoparticles as a novel adsorbent for thepreconcentration of mercury species prior to

cold vapor-atomic fluorescence spectrometric determination, *Anal. Sci.* 32 (2016) 867–873.

[26] S. Zhu, B. Chen, M. He, T. Huang, B. Hu, Speciation of mercury in water and fish samples by HPLC-ICP-MS after magnetic solid phase extraction, *Talanta* 171 (2017) 213-219.

[27] H. Cheng, X. Chen, L. Shen, Y. Wang, Z. Xu, J. Liu, Ion-pairing reversed-phase chromatography coupled to inductively coupled plasma mass spectrometry as a tool to determine mercurial species in freshwater fish, *J. Chromatogr. A* 1531 (2018) 104-111.

[28] M.L. Carvalho, M. Manso, s. Pessanha, A. Guilherme and FR. Ferreira, Quantification of mercury in XVIII century books by Energy Dispersive X-Ray Fluorescence (EDXRF), *J. Cult. Herit* 10 (2009) 435-438.

[29] A. Bernaus, X. Gaona, J.M. Esbri, P. Higueras, G. Falkenberg and M. Valiente, Microprobe techniques for speciation analysis and geochemical characterization of mine environments: the mercury district of Almadén in Spain, *Environ. Sci. Tech.* 40 (2006) 4090-4095.

[30] W. Stöber, A. Fink and E. Bohn, Controlled growth of monodisperse silica spheres in the micron size range, *J. Colloid. interf.* Sci 26 (1968) 62-69.

[31] J.S. Suleiman, B. Hu, H. Peng and C. Huang, Separation/preconcentration of trace amounts of Cr, Cu and Pb in environmental samples by magnetic solid-phase extraction with Bismuthiol-II-immobilized magnetic nanoparticles and their determination by ICP-OES, *Talanta* 77 (2009) 1579-1583.

[32] H. Tavallali, H. Shafiekhani, M.K. Rofouei and M. Payehghadr, A new triazene ligand immobilized on triacetylcellulose membrane for selective determination of mercury ion, *J. Braz. Chem. Soc* 25 (2014) 861- 866.